

Increase of Metal Accumulation in Plants Grown on Biochar – Biochar Ecotoxicity for Germinating Seeds

P. Soudek, Š. Petrová and T. Vaněk

Abstract—The effect of toxic metals on seed germination was studied in 3 cultivars of sorghum (*Sorghum bicolor* L.). Toxicity of cadmium, copper, and lead at five different concentrations (0.05–5 mM) was tested by standard ecotoxicity test. Three different biochars as supplement was used for the test. Root length was measured after 72 h of incubation. Elongation inhibition and EC₅₀ value were calculated. The results showed that beech tree biochar was the most efficient to reduce the toxicity of tested heavy metal. It contained the highest PAHs content, lowest content of measured heavy metals and the lowest pH value from tested biochars.

Index Terms—Biochar, heavy metals, seeds germination, sorghum.

I. INTRODUCTION

The area of land contaminated with heavy metals has increased during the last century due to mining, smelting, manufacturing, urban and other industrial activities. Nowadays, heavy metals have a serious ecological impact due to their toxicity and accumulation. These elements can be leached into the surface water or groundwater, taken up by plants, released as gases into the atmosphere, or bound semi-permanently by soil components such as clay or organic matter, and later affect human health [1], [2].

Heavy metals represent high stress factor for environment and for human. Their toxicities are expressed in many different effects and they can be cause of indigestions, different dermatitis, changes in blood count, damage of fundamental organs (brain, liver, and kidney), cancerous processes etc. For majority of metal cations the strong binding with –SH, –COOH and –NH₂ groups are characteristic [3]. Therefore more important is the lowering of toxic metals on the natural level.

Root growth inhibition as a parameter of heavy metal toxicity for plants was tested by many authors. Reference [4] tested responses of root growth inhibition and photosynthetic pigments production of *Sinapis alba* on heavy metal treatment (Cu²⁺, Ni²⁺, Mn²⁺, MoO₄²⁻, VO₄³⁻). The most toxic for root growth were Cu²⁺ and MoO₄²⁻. In the next paper [5] and [6] were investigated metal-metal interactions in accumulation of same metal treatment as above. Lead uptake and effect on seed germination and plant growth in Pb

hyperaccumulator *Brassica pekinensis* Rupr. was reported [7]. They found 56.7 % decreasing of germination rate at the concentration 1000 µg of Pb per mL and also decreasing of root length. The inhibitory effects of Cd, Cu, Zn, Pb and Fe on root elongation, contents of photosynthetic pigments, and metal accumulation in the roots and shoots of *S. alba* were tested [8]. Metals were arranged on the basis of growth inhibition in order Cu > Cd > Fe = Zn > Pb. Reference [9] reported stimulation effect of 0.005 mg/L dose of tested heavy metals (Cd²⁺, Cr⁶⁺, Cu²⁺, Ni²⁺ and Zn²⁺) on root and shoot elongation of alfalfa plant (cv. Malone). The dose of 0.04 mg/L of Cd²⁺, Cr⁶⁺, Cu²⁺ and Ni²⁺ significantly reduced the ability of the seed to germinate and growth in the medium with heavy metal treatment. Effect of metals on seed germination, root elongation and coleoptile and hypocotyl growth in *Triticum aestivum* and *Cucumis sativus* were also studied [10]. Mercury was determined to be the most inhibited metal under these parameters. This metal caused a complete inhibition of germination in wheat and cucumber seeds at certain concentrations: ≥ 1.5 mM in cucumber and at 1.7 mM in wheat.

The aim of the present work was to evaluate the effect of toxic metals in presence or absence of biochar on the germination of sorghum (*Sorghum bicolor* L.) seeds to select resistant cultivars for phytoremediation purposes.

II. MATERIAL AND METHODS

A. Plant Material and Chemicals

Three cultivars of *S. bicolor* L. (Expres, Honey Graze and Nutri Honey) seeds (obtained from SEED SERVICE s.r.o., Czech Republic) were used for the germination test. Heavy-metal ions (Cd²⁺, Cu²⁺, and Pb²⁺) were obtained from salts Cd(NO₃)₂ × H₂O, Cu(NO₃)₂ × 3H₂O, and Pb(NO₃)₂. Tested concentrations of each metal were 0.05, 0.1, 0.5, 1 and 5 mM. The biochar was used from beech tree (EKOGRILL Ltd., Czech Republic), rice husk or bamboo (Dr. Jing Song, Nanjing, China). All substances were dissolved in distilled water containing 2 mM CaCl₂ × 2H₂O, 0.5 mM MgSO₄ × 7H₂O, 0.8 mM NaHCO₃, and 0.08 mM KCl (according to ČSN EN ISO 7346) (all chemicals were from Penta [http://www.pentachemicals.eu]). The pH was adjusted to 7.6 by addition of 0.1 M solution of NaOH.

B. Semichronic Toxicity Test

The seeds were placed in plastic dishes (10-cm diameter) with a layer of a filter article on the bottom. To the half of Petri dishes was weighed 0.5 g of biochar. Seventeen seeds were equally placed into each dish on the surface of filter article, and 5 mL tested aqueous solution with heavy metal

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was added. Each treatment had four replicates. The exposure lasted for 72 h in the dark at 25°C. Root lengths were measured, and inhibition values of root elongation were calculated according to the following equation (1):

$$I = (Dc - Dt) / Dc \quad (1)$$

where I is the inhibition of root elongation in %; Dc is the average length of root under control conditions (i.e., without heavy metal treatment) [mm]; and Dt is the average length of root grown under the tested metal concentration [mm].

C. EC_{50} Calculation

EC_{50} is an effective concentration at which 50% of tested organisms have a significant response to the tested compound. For the purpose of calculation, nonlinear regression with the lower and upper maximum (0 and 100, respectively) was used. Data were processed by software GraphPad Prism (GraphPad, San Diego, CA), and the output was to MS Excel. The NOEC value is the highest tested concentration of a toxic compound at which there is no significant unfavorable impact on tested organisms compared with controls.

D. Heavy Metal Determination

Biochars were ground to a powder and samples (Ca 0.125 g DW) were digested in 5 mL mixture $HClO_4/HNO_3$ (15/85%, v/v) in digestion glass tubes overnight. Digestion was completed by gradual increase of temperature from 60 to 195 °C according reference [11]. Digestion protocol was as follows: 3 h, 60 °C; 1 h, 100 °C; 1 h, 120 °C; 3 h, 195 °C. After cooling, 2.5 mL 20% HCl was added, whirl mixed and warmed to 80 °C for 1 h. The final volume was brought to 10 mL accurately. Heavy metal content was measured by AAS (SensAA, GBS, Australia).

E. pH Measurement

The pH was determined on a 1:5 biochar / 0.01M $CaCl_2$ or biochar / distillate water suspension after 30 minutes of settling. Values of pH were determined by an UltraBasic pH Meter (Denver Instrument Company, USA).

F. PAHs Extraction and Determination

Biochar (4 g) together with anhydrous sodium sulfate (1 g) were placed in the Erlenmeyer flasks. They were extracted with 40 mL dichloromethane at 25 °C for 24 h.

The dichloromethane solution was evaporated to almost dryness using rotary evaporator (400 mbar, 70 rpm, water bath temperature 35 °C) and then dissolved in 2 mL cyclohexane. Wherein 1 mL cyclohexane was cleaned up by silica gel column and eluted with acetone / hexane (1:1, v/v). About 4 mL eluent was collected and subsequently evaporated to dryness under nitrogen gas. The PAHs were dissolved in 1 mL dichloromethane before HPLC analysis.

G. Statistics

Statistical analysis was performed based on STATISTICA (StatSoft, Tulsa, OK) software.

III. RESULTS AND DISCUSSION

Biochar is, due to its ability of the sorption characteristics

to reduce mobility and phytotoxicity certain organic and inorganic pollutants in the soil, able to eliminate limitations of phytoremediation. But the amount of information that relates to the application of biochar in order to improve efficiency of phytoextraction and phytostabilization is currently very limited. The main aim of this study is therefore to test biochar obtained from the pyrolysis of different biological sources on toxic residues and possible influence of biochar on germinated plants in presence of heavy metals.

TABLE I: METAL CONTENT IN BIOCHAR SAMPLES AND pH

		Bamboo	Rice husk	Beech tree
Metal [$\mu\text{g/g}$]	Cd	n.d.	n.d.	n.d.
	Zn	8.29	2.41	0.45
	Cu	0.61	0.65	0.32
	Fe	8.65	39.68	7.22
	Pb	n.d.	n.d.	n.d.
pH	H_2O	9.50	10.00	8.69
	$CaCl_2$	9.19	9.36	8.21

TABLE II: PAH CONTENT IN BIOCHAR SAMPLES

PAHs [ng/g]	Bamboo	Rice husk	Beech tree
Naphthalene	17.71	24.13	240.86
Acenaphthylene	0.00	0.00	0.00
Acenaphthene	0.00	0.00	8.59
Fluorene	24.62	62.42	58.87
Phenanthrene	34.25	199.71	211.04
Anthracene	6.17	37.71	22.97
Fluoranthene	3.20	6.84	63.82
Pyrene	40.26	17.92	65.80
Benzo(a)anthracene	0.37	5.52	32.75
Chrysene	0.52	1.90	20.28
Benzo(b)fluoranthene	0.15	0.08	2.13
Benzo(k)fluoranthene	0.00	0.41	1.59
Benzo(a)pyrene	0.75	8.19	7.41
Dibenzo(a,h)anthracene	1.10	2.98	0.91
Benzo(g,h,i)perylene	6.46	7.78	3.75
Indeno(1,2,3-cd)pyrene	0.00	0.00	0.00
Summa PAH	135.58	375.60	740.77

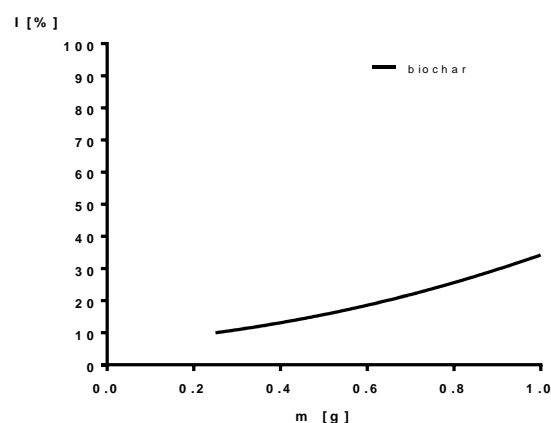


Fig. 1. Inhibition of seeds germination depends on biochar content.

Biochar pH was mildly alkaline and pH values were depended on different feedstock materials (Table I). The beech tree biochar contain the highest amount of PAHs (Table II). The most represented in the measured PAHs were naphthalene and phenanthrene. Fluoranthene, pyrene and fluorine were also significantly represented. Reference [12]

published that the naphthalene was the most abundant individual PAH compound in all of the biochar matrices assessed. Also anthracene, fluoranthene and pyrene were the most often found compounds present in the biochar matrices. The amount and content of PAHS in biochar is strongly depending on temperature of pyrolyzation. From the measured heavy metals cadmium and lead were not detected. Iron was significantly increased in biochar from rice husks (Table I).

TABLE III: EC₅₀ VALUES

Cultivar	Metal	Bamboo	Rice husk	Beech tree	Control
Expres	Cd	4.816	2.763	6.498	0.877
	Cu	2.837	2.379	9.144	2.054
	Pb	2.530	1.170	7.214	7.209
Honey Graze	Cd	12.03	6.141	18.34	0.929
	Cu	22.51	18.39	25.45	2.056
	Pb	7.348	5.513	8.450	6.032
Nutri Honey	Cd	6.196	3.616	23.23	0.977
	Cu	5.886	2.290	7.201	2.033
	Pb	1.097	3.889	7.221	5.884

The toxicity of biochar slightly increased with the amount of biochar in the medium (Fig. 1). The IC₅₀ value was calculated on 20.8 g of biochar.

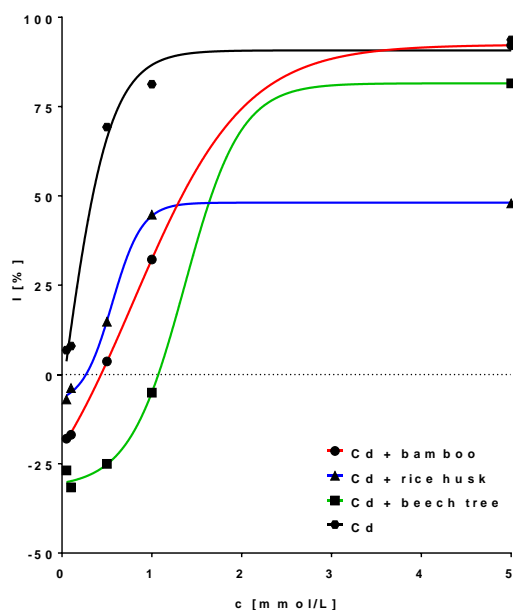


Fig. 2. Inhibition of sorghum seeds (cv. Nutri Honey) germination by cadmium in presence of different type of biochar.

Cadmium ion toxicity showed significant differences among the tested cultivars and especially between tested biochars. Beech tree biochar significantly decreased the toxicity of cadmium ions. Table III contained a list of EC₅₀ values in wide range (from 2.76 to 23.23 mM) for the biochars compare to close range (from 0.887 to 0.977 mM) for the tested cultivars without biochar addition. In general, cadmium toxicity was high and biochar addition decreased the toxicity ten times (Fig. 2). Cadmium was probably adsorbed on the surface of particles of biochar and thus reduced the actual concentration of cadmium. The reduced availability was described [13]. In the presence of biochar

reduced cadmium concentration in pore water more than ten times was detected.

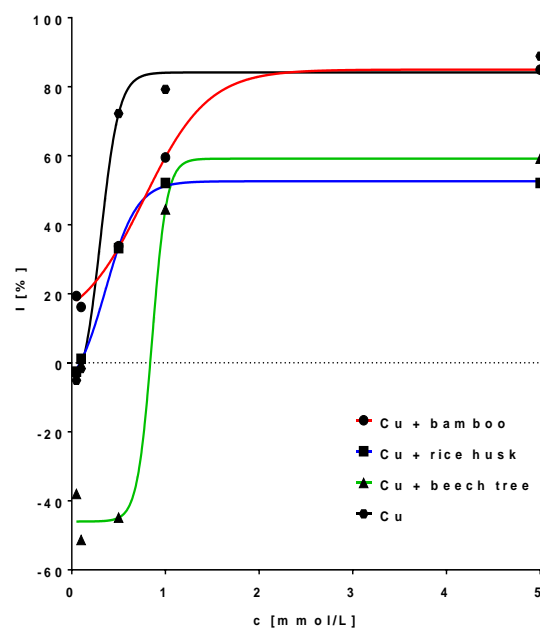


Fig. 3. Inhibition of sorghum seeds (cv. Nutri Honey) germination by copper in presence of different type of biochar.

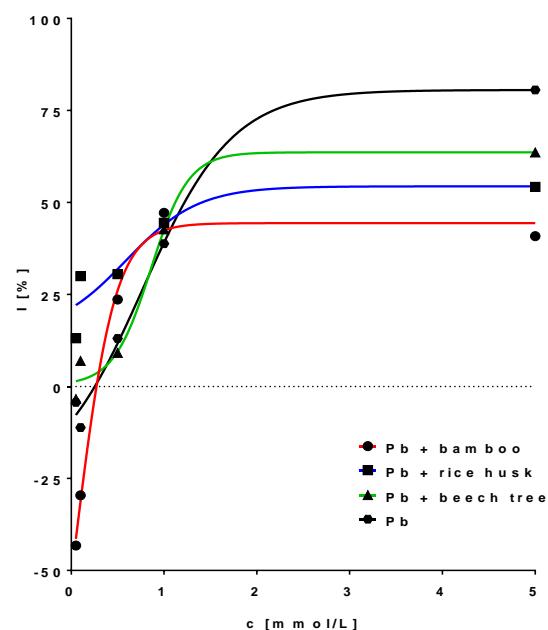


Fig. 4. Inhibition of sorghum seeds (cv. Nutri Honey) germination by lead in presence of different type of biochar.

In the case of copper, a wide range of EC₅₀ values was observed for biochar addition and very close range for copper toxicity for different cultivars without biochar (Table III). Our results showed relatively low sorghum sensitivity to the presence of copper and biochar together, with the differences between cultivars being rather large. The beech tree biochar was the most efficient in copper toxicity reduction and even at low concentrations copper stimulated seeds germination. For other biochar types we obtained reduction toxicity in case of cultivar Honey Graze (Fig. 3). For cultivars Nutri Honey and especially for cultivar Expres the EC₅₀ was close to EC₅₀ values for copper without biochar. As was published in reference [13], copper concentrations in pore water increased

after biochar addition and concentrations were significantly greater than in the control soil. Increased copper concentrations in pore water were associated with elevated concentrations of soluble organic carbon (DOC) from biochar. Associations between copper and DOC were previously found [14].

Slight stimulation of root growth was observed at a lead concentration 0.01 mM with bamboo biochar addition. The EC₅₀ values for lead ion in presence of biochar were found to be in the range of 1.10 to 8.45 mM. Cultivars showed a higher sensitivity to lead in presence of biochar than in its absence (Table III and Fig. 4). Lead is well known for its creation of insoluble compounds (i.e. with sulphate). Addition of biochar may cause better availability of lead due to an adsorption of lead ions to biochar particles. It was also reported that inactivated plant biochars can inadvertently increase dissolved lead and copper concentrations of sandy, low TOC soils and can be used to stabilize other contaminants [15]. These can explain the relatively high lead toxicity in presence of biochar.

IV. CONCLUSION

Our results showed a high diversity in the response of sorghum cultivars to the presence of toxic metals. We found wide differences in toxicity for cultivars in the case of Pb²⁺, and high influence of biochar presence on heavy metal toxicity. The differences between types of biochar were also found. Differences determined between biochars were mainly in PAHs content and pH. Our results proved the toxicity test is a useful tool for the testing of biochars suitable for phytoremediation and soil restoration purposes.

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